Determination of Selenium in Soils by Cathodic Stripping Voltammetry after Separation as Gaseous Selenium Tetrabromide

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A procedure for soils has been developed that uses a nitric/sulfuric acids digestion to solubilize total selenium and a distillation treatment to release selenite and selenate in the form of gaseous SeBr₄. The advantage of distillation treatment procedure is that analyte is separated from the interfering matrix and simultaneously converted to electroactive form Se(IV). Optimum procedures for decomposition of the soil samples and for separation of low contents of Se from the interfering components were investigated. This procedure was used for determination of extractable forms of Se as well. The selenium(IV) was determined by differential pulse cathodic stripping voltammetry (DPCSV) with hanging mercury drop electrode in the presence of added copper(II). Concentration of Se was evaluated by the method of standard additions on the basis of the signal arising on reduction of Cu(I) in the deposited copper(I) selenide. The developed procedure is simple as to the number of operations preceding the determination itself and reliable for determination of very low contents of Se in various types of soils and soil extracts as evidenced by low values of relative standard deviations (0.8—4.9 %). The detection limit for total selenium is $w = 5 \times 10^{-9}$ using a sample size of 0.3 g.

Selenium is an important metalloid with industrial, environmental, biological, and toxicological significance. Recent increased awareness of the biological role of Se has produced a revival of interest in methods for the determination of Se and its speciation at trace concentration. Measurement of Se in environmental samples is important as at low concentrations it is an essential trace element whereas at higher concentrations toxic effects occur. The narrow concentration range between deficient and toxic level of selenium requires precise knowledge of the selenium content in environment.

The content of total selenium in soils lies in general below $w = 1 \times 10^{-6}$ and seldom reaches higher values [1]. The low content of Se in soils demands a high sensitivity of the analytical method used for its determination as well as sample handling procedures. Among the methods which can be applied for the determination of traces of Se molecular fluorescence spectrometry, gas chromatography with electron capture detection, spectrophotometry and atomic spectrometry together with hydride generation are usefully employed for the routine laboratory determination of Se in environmental samples [2—4]. From electroanalytical methods stripping voltammetry is a suitable alternative to the methods mentioned above for the determination of low contents of Se. The advantage of this method is its simplicity, excellent sensitivity as well as inexpensive instrumentation.

Cathodic stripping voltammetry (CSV) with hanging mercury drop electrode (HMDE) is the most often used electroanalytical method for the determination of traces of Se and it was applied to analysis of soil samples, too [1, 5-7]. There is evaluated the signal arising on cathodic stripping of the deposited Se either in the form of Cu₂Se or HgSe. Under interference-free conditions, CSV can easily detect nanogram levels of Se in a sample with good accuracy and precision. However, in real sample, interferences from other elements can greatly limit the sensitivity of Se determination by CSV. Therefore, voltammetric determination requires separation of Se from the soil matrix. Selenium can be separated either directly from matrix by combustion of samples in oxygen [4, 5] or from digests using e.g. distillation in the form of SeCl₄, or SeBr₄ [4, 8, 9], coprecipitation with hydroxides and hydrated oxides and extraction of selenium(IV) halides into organic solvents [4, 8]. We have coped with this problem using distillation of Se in the form of gaseous selenium(IV) bromide prior to voltammetric determination. The advantage of this separation method lies in that reduction of the inactive Se(VI) to active Se(IV) takes place simultaneously with separation of Se from interfering matrix components.

EXPERIMENTAL

Hydrochloric, nitric, and sulfuric acids were Merck, suprapur. All other chemicals used were of anal. grade from Lachema, Fluka and were not further purified. The water used had been deionized and further purified in Labconco Water Pro PS system to "18 M Ω quality"

The Se(IV) standard stock solution $(1 \times 10^{-2} \text{ mol} \text{ dm}^{-3})$ was prepared by dissolving a measured quantity of selenium dioxide (BDH standard) in 0.1 M-HCl. The stock solution of Se(VI) $(c = 1 \times 10^{-4} \text{ mol} \text{ dm}^{-3})$ was prepared by oxidation of Se(IV) stock solution by addition of chlorine solution and by appropriate dilution. Working standard solutions of lower concentrations 2×10^{-6} to 1×10^{-5} mol dm⁻³ Se(IV) or Se(VI) were prepared before analysis by dilution of the stock solutions.

Analyzed were soil samples from Slovakia, taken according to usual procedure in the Agricultural Research Institute (Bratislava) (designed by N) and standard materials ECH No.12-1-03 and EOP No. 12-1-02 (light ashes from thermal power stations) (Institute of Radioecology and Utilization of Nuclear Technique, Košice). The soil samples were dried, homogenized and sieved to 0.2 mm.

The polarographic Analyzer PA 4 with X—Y recorder 4103 and Static Mercury Drop Electrode SMDE 1 (Laboratorní přístroje, Prague) were used for FSDPCSV measurements. The electrochemical cell was equipped with Ag/AgCl reference electrode (saturated KCl), the platinum auxiliary electrode and the working electrode which was used in the HMDE mode. Voltammetric measurements were made under the following conditions: deposition potential -0.35 V, deposition time 120 s, drop size 160 ms (2.3 mg), pulse amplitude -50 mV, scan rate 10 mV s⁻¹, current sensitivity 2 to 5 nA cm⁻¹, time constant of memory 100 ms.

Deaeration of analyzed solutions was performed with nitrogen which had passed through an acidic vanadium(II) scrubber solution in contact with zinc amalgam.

Decomposition of Samples

Procedure A

To the sample (0.3-0.4 g) in teffon crucible concentrated HNO₃ (4 cm³) and H₂O₂ (0.5 cm³) were added. The crucible was closed and the mixture was heated on a sand bath at 110-115 °C for at least 6 h until a feebly yellow solution resulted. If the diges-

Procedure B

Procedure *B* is identical to procedure *A* except that decomposition mixture H_2SO_4 —HNO₃ ($\varphi_r = 1$ 1) (4 cm³) was applied. After decomposition the mixture was carefully evaporated until dense fumes of SO₃ appeared.

Procedure C

To the sample (0.3-0.4 g) in teflon crucible concentrated H_2SO_4 (2 cm³) was added, followed by gradual addition of concentrated HF (2 cm³) in small portions. The crucible was closed and the mixture was heated on a sand bath at 110-115 °C for 6 h. Then the mixture was cooled down to ambient temperature, concentrated HNO₃ (2 cm³) was added and the mixture was heated again. After complete decomposition solution was carefully evaporated until dense fumes of SO₃ appeared.

Distillation of Se

The soil digest was quantitatively transferred into distillation flask with 10 cm³ of 10.5 M-H₂SO₄ and 200 mg of KBr and 10 mg of KBrO₃ were added. Distillation of SeBr₄ was performed in slow stream of nitrogen (100—120 cm³ min⁻¹) at 250—260 °C for 30 min using a simple glass apparatus. Gaseous SeBr₄ was collected in 5 cm³ of 0.3 M-HCl and after dilution to 10 cm³ the resulting solution for CSV measurement was obtained.

Extraction of Se from Soil Samples

Soil sample (2 g) was shaken with 20 cm³ of 2 M-HNO₃ (1 M-NaOH) for 6 h in polyethylene flask. Then soil extract was filtered (Filtrak 390), evaporated on a sand bath to near dryness and digested with HNO₃ (2 cm³) and HClO₄ (0.5 cm³). If the digestion of organic matter was incomplete and the final solution was not clear yellow, the acid mixture was added and heated again. This digest was elaborated by distillation as it was mentioned above.

Voltammetric Determination of Se(IV)

Distillate $(0.5-2 \text{ cm}^3)$ was transferred into the polarographic cell and diluted to 10 cm^3 with the solution of 0.1 mol dm^{-3} HCl. Then 20 mm^3 of CuCl₂ ($c = 1 \times 10^{-2} \text{ mol dm}^{-3}$) was added to measured solution. The solution was deaerated for 5 min by bubbling nitrogen saturated with water vapour. After deaeration of the sample the Se(IV) peak was recorded under the conditions mentioned above in the potential range of -0.35 to -0.75 V by cathodic stripping voltammetry with fast scan differential pulse record of the stripping step (FSDPCSV). For determination of Se(IV) the method of standard additions was used. It was realized by three additions (10 to 50 mm³) of Se(IV) standard solutions ($c = 2 \times 10^{-6}$ to 1×10^{-5} mol dm⁻³). The concentration of Se(IV) was evaluated using the peak height.

RESULTS AND DISCUSSION

The optimalization of chemical and instrumental parameters for CSV determination of Se(IV) was realized in the previous work [10]. Deposition of Se on HMDE in the form of intermetallic compound Cu₂Se at a large excess of Cu(II) over Se(IV) has shown to be more suitable for analytical applications than deposition of HgSe and it was exploited in all measurements. In the presence of Cu(II) detection limit was lower, linear range of calibration curve was prolonged and the influence of interfering metal ions and splitting of peaks was suppressed. Concentration of Se(IV) was evaluated on the basis of the signal arising on reduction of Cu(I) in the deposited copper(I) selenide during the cathodic polarization of the HMDE. The results presented made it possible to specify the optimum conditions for the CSV determination of trace amounts of Se(IV), i.e. supporting electrolyte 0.01 to 0.5 M-HCl, $c(Cu(II)) = 1 - 2 \times 10^{-5} \text{ mol dm}^{-3}$, deposition potential -0.3 to -0.35 V, deposition time 120 s, scan rate $10 - 20 \text{ mV s}^{-1}$. The instrumental response under these experimental conditions was linear up to concentration of 5 $\mu g dm^{-3}$ (correlation coefficient = 0.999). The detection limit for direct DPCSV under adopted conditions is 0.15 $\mu g dm^{-3}$.

In the environment Se is present in inorganic and organic forms. Inorganic selenium can exist in several oxidation states: selenide, elemental selenium, selenite, and selenate. Organic selenium is generally present in the form of selenoamino acids and their derivatives or more or less volatile methylated species. First the procedure has been optimized for the determination of total selenium in soils. Most procedures for the determination of total selenium employ wet oxidation with acid mixtures to solubilize selenium. Comparative tests of digestion method employing various acid mixtures such as HClO₄ and HNO₃, HNO₃ and H₂SO₄, HNO₃, HF, and H₂SO₄, HNO₃ and H₂O₂ were conducted to ensure the total recovery of Se from the soil. Data obtained showed that digestion procedures proposed gave comparable concentrations for selenium in soil samples. Repeated digestion of soil samples with additional acids did not change the results of analyses. Despite of volatility of SeF_4 and SeF_6 no less of Se was found out if digestions, which included hydrofluoric acid were applied. The possible volatilization of Se from the system HF-HClO₄ under con-



Fig. 1. Voltammogram for Se(IV) recorded by FSDPCSV directly in soil digest (a) and after distillation pretreatment (b). 1. Signal of Se(IV) in sample ECH; 2, 3. after standard additions of 20 mm³, 40 mm³ of solution $(c(\text{Se(IV)}) = 1 \times 10^{-5} \text{ mol dm}^{-3}).$

ditions that are representative for the dissolution of silicates was investigated by Bajo [11]. No losses of Se by volatilization during the decomposition step were observed, which is in accord with our results.

After oxidative digestion all selenium is present in inorganic form as Se(IV) and Se(VI). Voltammetric methods are sensitive only for Se(IV) as Se(VI) cannot be reduced electrochemically. Therefore the reduction pretreatment prior to the voltammetric determination of total Se is desirable. The reduction of Se(VI) to Se(IV) with hydrochloric acid [1, 5, 7, 10] is suitable method and it was used in this work. The digestion solution was acidified to 4 M-HCl and heated in boiling water bath for at least 20 min. As follows from Fig. 1a the high concentration of acids and relatively large amounts of interfering metal ions interfere with Se(IV) measurements by DPCSV and make the determination of very low content of Se in soil digest difficult. Only a very small, scarcely discernible stripping peak situated on a high and sloping background current was obtained directly from sample digest. This means that the analyte has to be separated from decomposition solution prior to its voltammetric determination by use of a suitable separation method. Several separation procedures for selenium have been described in the literature, including distillation of volatile compounds of Se [4, 8, 9], coprecipitation [4, 8], derivatization followed by extraction into a small volume of an organic solvent [12, 13], and solid phase extraction [14]. This separation step prior to determination improves the detection limit and increases the precision and accuracy of the results.

The high concentration of ions in soil digest limits the application of anion-exchange resins but the volatility of some Se compounds can be successfully used for separation of Se from the complex matrix by distillation. In this work the reaction of Se(IV) and Se(VI) with HBr has been utilized for separation of selenium from the soil matrix. These forms of Se by heating the samples with HBr-Br₂-H₂SO₄ [8] and NH₄Br—condensed phosphoric acid reagent [9] evolve gaseous selenium tetrabromide. The choice of suitable reagent for distillation method was the problem because NH₄Br and HBr were not available for us. Therefore the effect of various amounts of KBr (NaBr) on distillation recovery of selenium(IV) standard samples both in concentrated H₃PO₄ medium and 7.5-10 M-H₂SO₄ was investigated. Initial experiments demonstrated that full recoveries of Se were not achieved. The yields of Se varied from 30 to 70 % for 300 ng of selenium. Various kinds of absorption solutions for trapping of the distilled selenium tetrabromide were examined, too. Higher recovery was not obtained even in the absorbing solution of 0.3 M-HCl and 0.6 M-HClO₄ ($\varphi_r = 1$ 1), which was recommended by Terada et al. [9] as the most suitable. Quantitative recoveries were found when small amount of KBrO₃ was added. The bromine-bromide redox buffer solution prevented the reduction of Se(IV) to the elemental form. From the results it follows that 100-200 mg of KBr (NaBr) and 10 mg of KBrO3 seem appropriate. Under these conditions both Se(IV) and Se(VI) are converted into gaseous selenium tetrabromide. The distillation serves as the simultaneous reduction and separation step. Selenium has been completely distilled during 15—20 min when the temperature has reached about 250 °C. The recoveries of 79—315.8 ng of Se(IV) and Se(VI) heated with 10 mg of KBrO₃ and 200 mg of KBr both in 10.5 M-H₂SO₄ and concentrated H₃PO₄ are shown in Table 1. Selenium was recovered from these solutions within 94.1—101.4 % with relative standard deviations less than 5 % in each case. These results demonstrate the utility of the method for separation of Se at the nanogram level.

The overall pretreatment procedure proposed (digestion with HNO₃ and H₂O₂ (A), HNO₃ and H₂SO₄ (B), HNO₃, H₂SO₄, and HF (C) and Se distillation in the form of SeBr₄) was applied to the choice samples. Concentrated phosphoric acid etches glass walls of distillation flask at temperature about 250 °C. On the basis of this fact distillation only from H₂SO₄ medium was recommended for next analyses. Data summarized in Table 2 show that all digestion procedures give comparable values for selenium, which is in accord with previous experiences.

The adequacy of pretreatment by distillation follows from Fig. 1b. Distillation eliminates the interferences in the DPCSV method that are observed if soil digest is directly analyzed. The Se(IV) peaks are welldefined and adequate for quantification of selenium.

Reference materials ECH, EOP have not certified value of Se. Our results obtained by DPCSV show reasonable agreement with those of hydride generation/AAS (EOP $w = 0.39 \times 10^{-6}$, ECH $w = 2.07 \times 10^{-6}$).

The overall pretreatment procedure and the CSV technique for the determination of Se were evaluated by analyzing soil samples spiked with known amounts of Se(IV) and Se(VI) because reference materials with

m(Se, given)/ng		m(Se, found)/ng		$s_{ar{\mathbf{x}}}/\%$		Recovery/%		Number of
Se(IV)	Se(VI)	Se(IV)	Se(VI)	Se(IV)	Se(VI)	Se(IV)	Se(VI)	measurements
10.5 M-H ₂ SO ₄								
315.8		305.3 ± 12.6	a and a start of	1.7		96.7		6
79.8		75.8 ± 6.5		3.7		96.1		7
	315.8		310.2 ± 18.2		2.5		98.2	7
	157.9		156.1 ± 15.6		4.3		98.9	7
	79.0		77.3 ± 5.2		2.9		97.9	7
Concentrated H ₃ PO ₄						-199 Adj		
315.8		320.2 ± 14.4		1.9		101.4		9
157.9		149.8 ± 12.3		3.1		94.9		5
79.0		77.7 ± 1.6		0.8		98.3		6
	315.8		297.1 ± 22.9		3.3		94.1	8
	197.4		191.1 ± 8.1		1.6		96.8	5

Table 1. Determination of Se in Standard Samples after Distillation in the Form of SeBr4 from Various Media

Intervals of reliability and relative standard deviations were computed from the dispersion variance for the significance level 0.05.

Table 2. Comparison of Results Obtained by Different Digestion Procedures

Sample	w 10 ⁶	$s_{ar{\mathbf{x}}}/\%$	Digestion method	Number of measurements
	0.27 ± 0.01	0.9	A	5
N-2-PD	0.29 ± 0.01	1.3	В	5
	0.34 ± 0.03	3.4	C	5
	0.38 ± 0.01	1.1	Α	6
N-1-PD	0.35 ± 0.04	4.3	В	5
	0.40 ± 0.05	4.9	C	6
	0.24 ± 0.01	2.1	Α	5
N-5-ZV	0.22 ± 0.01	1.6	В	5
	0.26 ± 0.01	2.6	C	10
ECH	2.42 ± 0.12	2.2	В	10
	2.38 ± 0.09	1.6	C	5
EOP	0.52 ± 0.03	2.2	Α	5
	0.58 ± 0.01	0.8	В	5

Table 3. Results of Determination of Total Se in Selected Soil Samples

Sample	Horizon	Depth/cm	$w \cdot 10^6$	$s_{ar{\mathbf{x}}}/\%$	Number of measurements
N-1-TT	1	5—15	0.52 ± 0.03	2.4	10
	2	30—40	0.24 ± 0.03	4.9	10
	3	7080	0.09 ± 0.004	1.6	5
N-2-LM	1	10—20	0.26 ± 0.01	2.3	9
N-3-GA	1	5—15	0.28 ± 0.02	3.3	10
N-1-NR	1	10—20	0.31 ± 0.03	3.7	10
N-17-DS	1	10—20	0.76 ± 0.04	2.2	10
N-5-ZV	1	10—20	0.22 ± 0.01	1.6	5

certified value of Se were not available. The recoveries of the known additions of Se substantiate the accuracy of the procedure and these results demonstrate the utility of the method for determination of Se at the nanogram level in soils.

Data obtained showed that for the determination of selenium total decomposition of soil samples is not necessary. Therefore, only digestion procedure utilizing the H₂SO₄—HNO₃ mixture was applied to next analyses of soils of varying origin. The results of determination of total selenium are presented in Table 3. The Se content is within the range reported for soils and mostly less than the average ($w = 0.4 \times$ 10^{-6}) [15]. The Se concentrations were evaluated by the method of standard additions. Peak height measurement was used for the quantification of Se. The plots of peak height vs. added Se(IV) concentration showed good linearity with correlation coefficients r > $0.999 \ (n = 4)$. From the presented data it follows that the precision of determination at this low concentration level is good. The relative standard deviations of Se obtained from five to ten determinations of each sample are in the range 0.8-4.9 %.

The present study evaluates the suitability of treat-

ment procedures for the CSV determination of Se in soil samples. The main advantage of this treatment procedure is that analyte is separated from interfering matrix and simultaneously converted to electroactive form with HBr arising from KBr in the presence of H_2SO_4 . From the presented data it follows that the proposed digestions are adequate and distillation step prior to determination increases the precision and accuracy of the results. On the basis of these facts we propose distillation using the volatility of SeBr₄ as the suitable pretreatment procedure prior to the CSV determination of very low contents of Se in complex samples, such as soils.

The measurement of the total concentration of a trace element provides little information about the bioavailability and potential risk since the toxicity of elements and their environmental mobility are dependent on their chemical forms. To obtain information about the bioavailable fraction of elements a simple step extraction procedure is usually used. In the present work only acidic (2 M-HNO₃) and alkaline (1 M-NaOH) extractions were applied to soil samples. Results of Se for soil extracts are presented in Table 4. Application of the alkaline extraction of the soil

Table 4. The NaOH and HNO3 Extractable Se Contents of Some Selected Soils

Sample	NaOH	Extraction		HNO ₃ Extraction			
	$w(Se, extracted) 10^6$	$s_{ar{\mathbf{x}}}/\%$	Recovery/% ^a	$w(Se, extracted) \cdot 10^6$	$s_{ar{\mathbf{x}}}/\%$	Recovery/% ^a	
N-1-PD	0.25 ± 0.01	2.5	66.3	0.05 ± 0.004	3.9	12.7	
N-1-TT/1	0.29 ± 0.02	4.0	55.8	0.03 ± 0.002	3.7	6.0	
N-2-PD				0.04 ± 0.002	2.4	12.3	
N-17-DS	0.47 ± 0.03	2.9	61.4				

a) Results in Tables 2 and 3 were taken as 100 % recovery.

samples results in (61 ± 5) % reported for total Se. Selenium species solubilized by sodium hydroxide treatment are: selenite, selenate, and Se bound to organic matter (notably humic substances). The content of selenite and selenate in alkaline soil extracts can be estimated after the removal of organic matter from soil extracts using XAD resins. As follows from Table 4 HNO₃ extraction released from soils less than 13 % of selenium.

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